



Article

Evolution of SO₂ and NO_x Emissions from Several Large Combustion Plants in Europe during 2005–2015

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Abstract: The aim of this paper is to investigate the evolution of SO₂ and NO_x emissions of ten very large combustion plants (LCPs >500 MW) located in the European Union (EU) during 2005–2015. The evolution of NO_x and SO₂ emissions were analyzed against the EU Directives in force during 2005–2015. The investigation was performed using space-borne observations and estimated emissions collected from the EEA (European Environment Agency) inventory of air pollutant emissions. The power plants were chosen according to their capacity and emissions, located in various parts of Europe, to give an overall picture of atmospheric pollution with NO_x and SO₂ associated with the activity of very large LCPs in Europe. Satellite observations from OMI (Ozone Monitoring Instrument) are compared with calculated emissions in order to assess whether satellite observations can be used to monitor air quality, as a standard procedure, by governmental or nongovernmental institutions. Our results show that both space observations and estimated emissions of NO_x and SO₂ atmospheric content have a descending trend until 2010, complying with the EU Directives. The financial and economic crisis during 2007–2009 played an important role in reducing emissions.

Keywords: Large Combustion Plant (LCP); NO_x; SO₂; emissions; Ozone Monitoring Instrument

1. Introduction

Atmospheric pollution is one of the largest environmental health risks in Europe, severely affecting human health and causing more than 400,000 premature deaths each year [1]. More than 50% of the SO₂ and NO_x emissions in Europe are associated with the production and distribution of fossil fuels [2]. It is well known that the emissions from thermal processes of fossil fuels have negative effects on human health and the environment. High levels of NO₂ can lead to cardiovascular dysfunctions and respiratory problems such as cold, bronchitis, asthma, and lung cancer [3,4] while the effects of sulfur dioxide on human body consists of irritation of the airways, coughing, shortness of breath and a sensation of tightening around the chest [5]. Large quantities of SO₂ and NO₂ can lead to acid rains and disturbances in the functioning and structure of ecosystems, e.g., the acidification of soils and waters [6].

Large combustion plants (LCPs) are combustion plants with a total rated thermal input equal to or greater than 50 MW and have an important contribution to the air quality degradation due to anthropogenic pollutant emissions in the atmosphere [7]. EU legislation has set specific emission

limit values for NO_x, SO₂, and dust emissions from installations with a rated thermal input of ≥50 MW. LCPs use fossil fuels to produce thermal or electric energy, resulting in residues and waste products (including emissions) that affect the quality of all environmental components, with a specific and clear impact on the atmospheric composition. Large amounts of trace gases (NO_x, SO₂, CO₂) and particulate matter (PM) are released in the atmosphere when LCPs are operational.

The first EU policy related to emissions control dates from the 1980s. Between 2005 and 2015 the main legislation related to the power plants was governed by: the LCP Directive (Directive 2001/80/EC) [8], the NEC (National Emission Ceilings) Directive (2001/81/EC) [9], the IPPC Directive (Directive 2008/1/EC) [10] and the Industrial Emissions Directive (IED, Directive 2010/75/EU) [11]. The Directive 2001/80/EC set up limitations for sulfur dioxide (SO₂), nitrogen oxides (NO_x), and dust emissions from large combustion plants. The NEC Directive introduced upper limits of national emissions for five important air pollutants: nitrogen oxides (NO_x), non-methane volatile organic compounds (NMVOCs), sulfur dioxide (SO₂), ammonia (NH₃), and fine particulate matter (PM_{2.5}). The measures of the LCP Directive were binding from 2008 until December 2015. As of 2010, according to the NEC Directive, all the EU states are required to meet their emission ceilings. Also, since 2010 the IED Directive imposed stack emission control and thresholds for each trace gas or PM, in order to reduce emissions. Note that the LCP and NEC Directives refer to reduction plans for each EU country at a national level by specifying a maximum total amount of emissions per year, without imposing specific measures for a certain plant. The Directive 2010/75/EU (IED), which replaced Directive 2001/80/EC (LCPD), comes with more restrictive measures, by constraining emissions locally, at the point where these leave the installation. However, exemptions from regulations for many EU countries (the so-called “Transitional National Plan”) could prolong the implementation of the reduction measures until 30 June, 2020 [11]. The restrictive measures of the LCP and NEC Directives are less tight than the new IED Directive. Consequently, the EU countries were free to manage the emission ceiling by closing certain power plants or reducing the number of operating hours [8,9].

Here we focus on two trace gases, NO₂ and SO₂, which are released into the atmosphere during thermal processes of fossil fuels when LCPs are operational. One of the purposes of this paper is to test whether satellite measurements of NO₂ and SO₂ can be used as proxies for NO₂ or SO₂ emissions over different areas of the Earth's surface, thus replacing or supplementing the estimated emission calculations.

Satellite-based observations of atmospheric parameters have many applications, ranging from climate change monitoring to trace gas observations [12,13]. Moreover, space observations can provide self-consistent information about the evolution of NO₂ and SO₂ on a continuous-time basis, based on daily global coverage [13]. Spectroscopic measurements of NO₂ and SO₂ using UV-Vis DOAS space observations have been available since July 1995, when the Global Ozone Monitoring Experiment (GOME-1) was launched into space onboard ESA's 2nd European Remote Sensing Satellite (ERS-2) [14,15]. Space-based DOAS observations of tropospheric NO₂ and SO₂ have, since then, become a very useful tool for monitoring of emissions from different sources at a global or regional level [16–18]. One such tool is the Ozone Monitoring Instrument (OMI), which is a space UV-Vis spectrometer based onboard of the AURA satellite, used for NO₂, SO₂ and other trace gas observations [19]. Space observation can be used for estimation of the quantity of NO₂ and SO₂ from various sectors of activity, including energy production [20–23].

Trends of regional NO₂ or SO₂ reported by satellite observations are different. For instance, a decrease of SO₂ and NO₂ emissions are reported for China [24,25], USA [26,27], and Europe [28], while increased emissions are observed for India [29]. Emissions from LCPs constitute a large proportion of total anthropogenic emissions. In 2015, LCP emissions of SO₂ and NO_x made up 44% and 14%, respectively, of total EU-28 emissions of these pollutants [2]. EU imposed specific emission limit values on emissions of NO_x, SO₂, and dust from plants with a rated thermal input equal to or greater than 50 MW. Also, an important role in the emission reduction was induced by the deindustrialization of many European countries; the most affected being the countries of Eastern Europe.

Figure 1 shows the evolution of the environmental performance of LCPs during 2004–2015 in the EU-28, expressed as implied emission factors for SO₂ and NO_x by fuel type. The implied emission factor is the ratio between emissions and fuel consumption. The implied emission factor for the NO_x and SO₂ pollutants decreased significantly between 2005 and 2015 for large combustion plants of different sizes [30]. However, according to the 2015 and 2017 indicator assessment [31,32] presented by the EEA, the emission reductions cannot be linked only to environmental policies implementation, but to other factors as well, e.g., broader economic and societal changes, economic conditions, international fuel prices, industry initiatives, etc. [30]. Singhal in 2019 presented a comprehensive study regarding the emissions reduction from LCPs in the European policies context; he concluded that the LCP Directive was an effective instrument in pollution abatement at the stack-level [33]. Also, Meyer and Pac (2017) [34] discussed the consequences of the LCP Directive over the 1585 EU's large combustion plants.

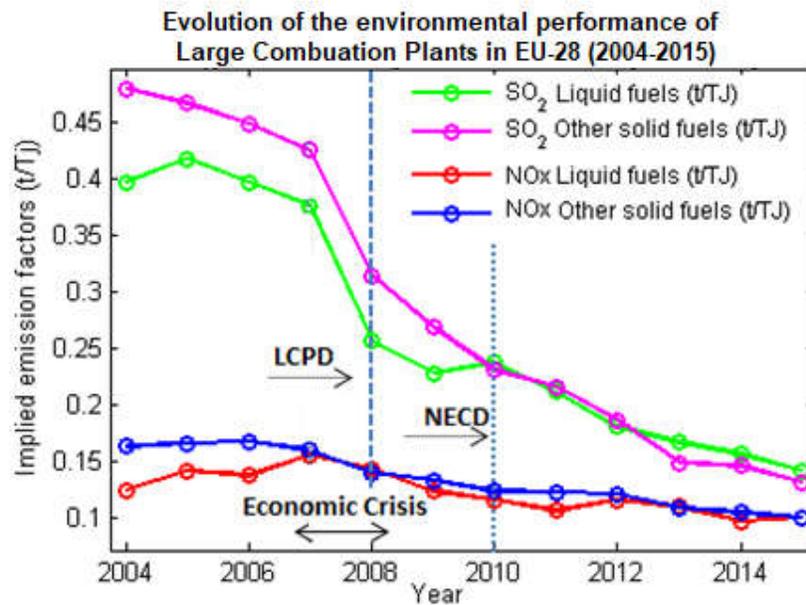


Figure 1. Evolution of the environmental performance of large combustion plants in the EU-28, expressed as implied emission factors for SO₂ and NO_x by fuel type.

The aim of this study is to analyze the evolution of NO₂ and SO₂ using space observations of the OMI instrument and EEA reported emissions for ten large power plants located in the EU, in order to assess the effect of EU standards and regulations implementation. This paper is organized as follows. Data and methodology are described in Section 2, results and discussions are presented in Section 3, while Section 4 is dedicated to conclusions.

2. Materials and Methods

Ten European power plants were selected, considering their capacity and quantity of NO_x and SO₂ emissions (<https://www.eea.europa.eu/data-and-maps/data/lcp-9>). The LCPs are located in various parts of the European continent (Figure 2) and are considered as very large combustion plants (>500 MW). Location (i.e., latitude and longitude) and the annual average of NO_x and SO₂ emissions level for the selected power plants are presented in Table 1.



Figure 2. The spatial distribution of LCPs; these are identified by the corresponding numbers in Table 1.

Table 1. Information for individual power plants (source: EEA *).

Power plant	Country	Code	Lat. (°)	Long. (°)	Average annual power (MWth)	Average annual SO ₂ emissions (T)	Average annual NO _x emissions (t)
1. TPP “Maritsa Iztok 2”	Bulgaria	BG	42.25	26.13	6743	232,084	16,211
TPP “Maritsa Iztok 3”			42.05	25.62			
Stara Zagora							
2. KW Jänschwalde, Peitz	Germany	DE	51.83	14.46	9144	21,438	19,218
3. PPC S.A.–Megalopoli I-IV	Greece	GR	37.41	22.10	2381	108,620	4543.5
4. Elektrownia “Kozienice” S.A.	Poland	PL1	51.66	21.46	7023.1	41,334	20,544
5. PGE Górnictwo i Energetyka Konwencjonalna S.A.–Oddział Elektrownia Bełchatów, Łódź Voivodeship	Poland	PL2	51.26	19.33	13170	80,789	40,217
6. S.C. Complexul Energetic Oltenia S.A., Rovinari 1-2	Romania	RO1	44.90	23.13	3512	60,166	11,937
7. S.C. Complexul Energetic Turceni S.A. 1-4	Romania	RO2	44.66	23.41	4734	72,824	14,526
8. CT LITORAL I-II, Carboneras-Almeria	Spain	SP1	36.97	−1.90	2737.3	14,929	10,453
9. Central térmica de Puentes de García	Spain	SP2	43.44	−7.86	3795.3	76,525	10,783

Rodríguez, La Coruña (CT AS PONTES I-II-III-IV)							
10. Drax Power Limited, Drax Power Station	United Kingdom	UK	53.73	−0.99	10145	26,094	42,954

* <https://www.eea.europa.eu/data-and-maps/data/lcp-9>.

The NO₂ and SO₂ satellite data were provided by the OMI space instrument as grid-averaged columnar amounts. OMI is a nadir-viewing UV-Vis spectrometer that measures atmospheric trace gases and aerosols. It provides daily global observations at a resolution of 13 km × 24 km. NO₂ was gathered from the Tropospheric Emission Monitoring Internet Service (TEMIS) database (<http://temis.nl/airpollution/NO2.html>). OMI monthly-mean tropospheric NO₂ columns were based on the Dutch OMI NO₂ version 2.0 product, which is a post-processing data set performed at Koninklijk Nederlands Meteorologisch Instituut (KNMI) [35]. We used the ESRI grid format with a cell size 0.125 degree. Note that the OMI sensor can provide only information about NO₂; in the case of intercomparisons, the NO_x emissions estimated from the ground will be expressed as NO₂ equivalent. The SO₂ data were obtained from the NASA Geospatial Interactive Online Visualization ANd aNalysis Infrastructure (Giovanni) interface, which is a remote-sensing and model data Web-based analysis and visualization system developed by the Goddard Earth Sciences Data and Information Services Center (GES DISC) [36]. We used the SO₂ Column Amount (Planetary Boundary Layer) OMSO₂e v003 [37] available on <https://giovanni.gsfc.nasa.gov/giovanni/>. This is a Level-3 Aura/OMI global SO₂ data product, based on grids (0.25-degree Latitude/Longitude grids) containing one observation of total column density of SO₂ in the planetary boundary, derived from an improved band residual difference algorithm (BRD) [38,39]. The NO₂ and SO₂ columnar amount, within a grid cell of 0.25 degrees centered on each LCP center, were considered for this study. Data regarding NO_x and SO₂ emissions between 2005 and 2015 are obtained from the European Environment Agency online database.

3. Results

This section presents the evolution of SO₂ and NO₂ emissions reduction as observed from space or derived from the ground in the context of main industrial emissions directives which governed the period 2005–2015, especially the LCP and NEC Directives. The implications and main drivers that could lead to decreases or increases in emissions are introduced in this section.

Figures 3 and 4 present maps of NO₂ and SO₂ tropospheric amounts observed by OMI, over Europe during 2005 and 2010. We show 2005 because this is the first full year of OMI measurements and 2010 because this was the milestone year of the NEC Directive. Hotspots can be clearly associated with the power plants, e.g., Greece, Bulgaria, and Romania. The NO₂ decrease over the selected power plant is visible from OMI (Figure 3) but the most important drop is visible in the case of SO₂ (Figure 4).

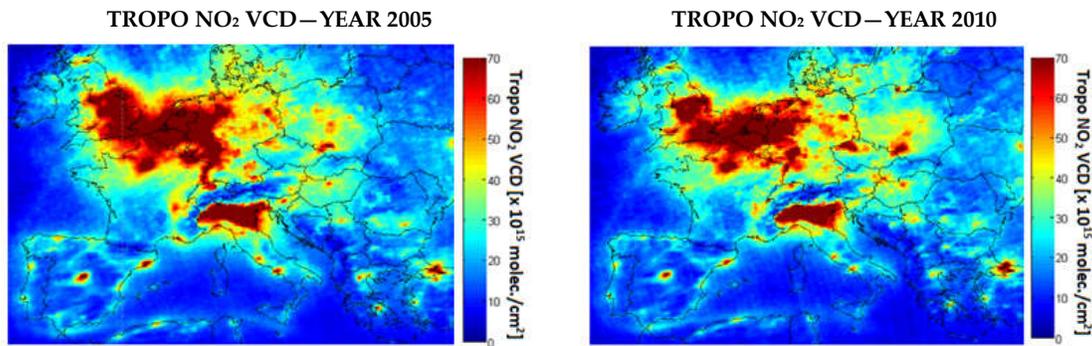


Figure 3. Annual tropospheric NO₂ VCD based on OMI observations for 2005 vs. 2010.

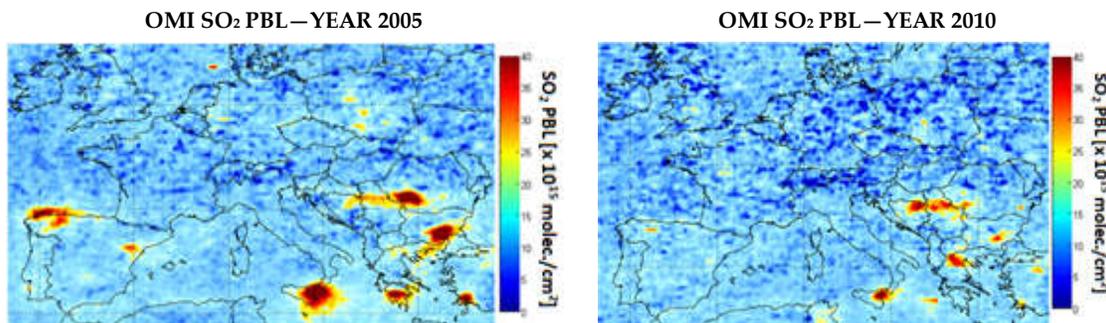


Figure 4. Annual tropospheric SO₂ PBL based on OMI observations for 2005 vs. 2010.

3.1. Nitrogen Oxides

Figure 5 shows the NO₂ evolution for each power plant, during 2005–2015, resulting from OMI measurements of the tropospheric NO₂VCDs and from estimated emissions. The latter are reported by each EU country to the EEA. Plots include the consumption of liquid and solid fuels per LCP. All measurements are normalized to their maximum. Expectedly, emission calculations follow roughly the trend of solid/liquid consumption, since the calculation of the former are based on the latter.

Almost all reported emissions of NO₂ showed a decreasing trend, except plants in BG and DE, while OMI space observations show an increasing trend for all, except plants in BG and GR. Note that one cannot expect a one-to-one correspondence between the reported emissions and space observations. On the one hand, differences between emissions and OMI stem from the fact that the satellite instrument sees a large area (13 km × 24 km), thus measurements contain background NO₂ and the NO₂ released by surrounding local/traffic sources. Secondly, satellite results are subject to various assumptions related to the atmospheric mass factor (AMF) calculations. The emissions, on the other hand, are based on calculations based on the quantity and nature of fossil fuel. The type and quality of fuel may influence the quantity of emissions, the correlation between the quantity of calculated emissions, and the nature of the fossil fuel (solid or liquid) which could give the main fuel used during the thermal processes. The discrepancies may also be due to the fact that OMI soundings are done at a specific time (i.e., the overpass time), which is associated with various phases of the diurnal variation of NO₂, depending on the geographical location of the station.

OMI measurements and estimated emissions correlate fairly well ($R > 0.5$) for a few power plants: BG, RO2, and SP1. However, for the other plants, there is practically no correlation since minima and maxima in the two-time series are even opposing (e.g., SP2, DE, and GR). A clear peak is seen in the NO₂ satellite time series around 2010–2011, when the winter was unusually cold in Europe [40], for all stations except stations GR, SP1, and the UK. The first are the southernmost ones,

which probably were less affected by the low temperatures, while the UK had a different climate compared to the continent. The increased NO₂ concentration may be the result of a combination of the increased request for heating and a higher lifetime of NO₂, of the order of days, when temperatures are low. An important characteristic of the NO₂ evolution, very well highlighted in OMI space observations, was the global financial and economic crisis during 2007–2009 [26,41–43]. Considering the influence of the financial and economic crisis (2007–2009) on the emissions, the end of recessions could correspond to the emissions increase in 2010 and after this year.

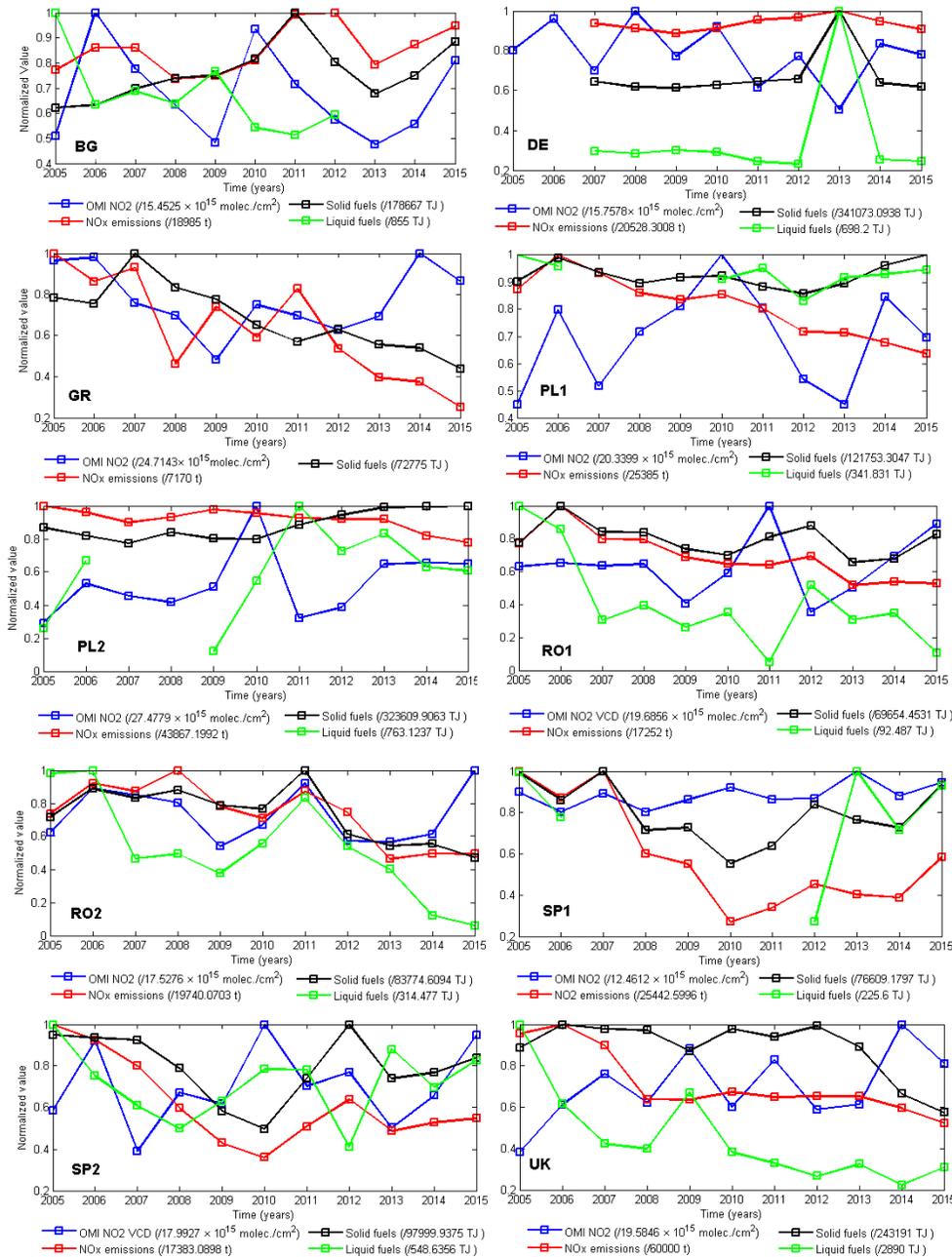


Figure 5. Time series of normalized NO₂ during 2005–2015 using data from satellite instruments (blue) and emissions (red) for each station, together with the reported solid fuel (black) and liquid fuel consumption (green).

Scatter plots of OMI measurements against the reported emission for NO₂ amount are shown in Figure 6, for all plants. Satellite measurements seemed to be a good indicator for emission variability only for some LCPs: EP1, EP2, BG, and RO2, for which correlation coefficients were positive ($R = 0.56, 0.52, 0.26, 0.32$). For other plants, the emissions varied in a completely different manner than the satellite measurements of tropoVCDs of NO₂ and vice versa. Large negative correlations between the two datasets were seen, e.g., in the UK ($R = -0.54$) or DE ($R = -0.67$). Indeed, as mentioned before, satellite data will inevitably contain also the background NO₂ resulting from all sources in an area of about 300 km². VCDs also respond to background conditions (higher or lower temperatures, wind effect, urban agglomeration), while emissions relate only to fuel consumption. For LCPs where OMI showed an increasing trend while the calculated emissions show the opposite, the explanation could be the fact that the calculated emissions from LCP do not include the nearby emissions, as urban and traffic emissions, or other industries; the latter are however included in space observations. This may be the case since Figure 5 shows that towards the end of the selected period, most emissions decreased, while tropospheric VCDs increased. However, the very large differences seen for other stations cannot be explained.

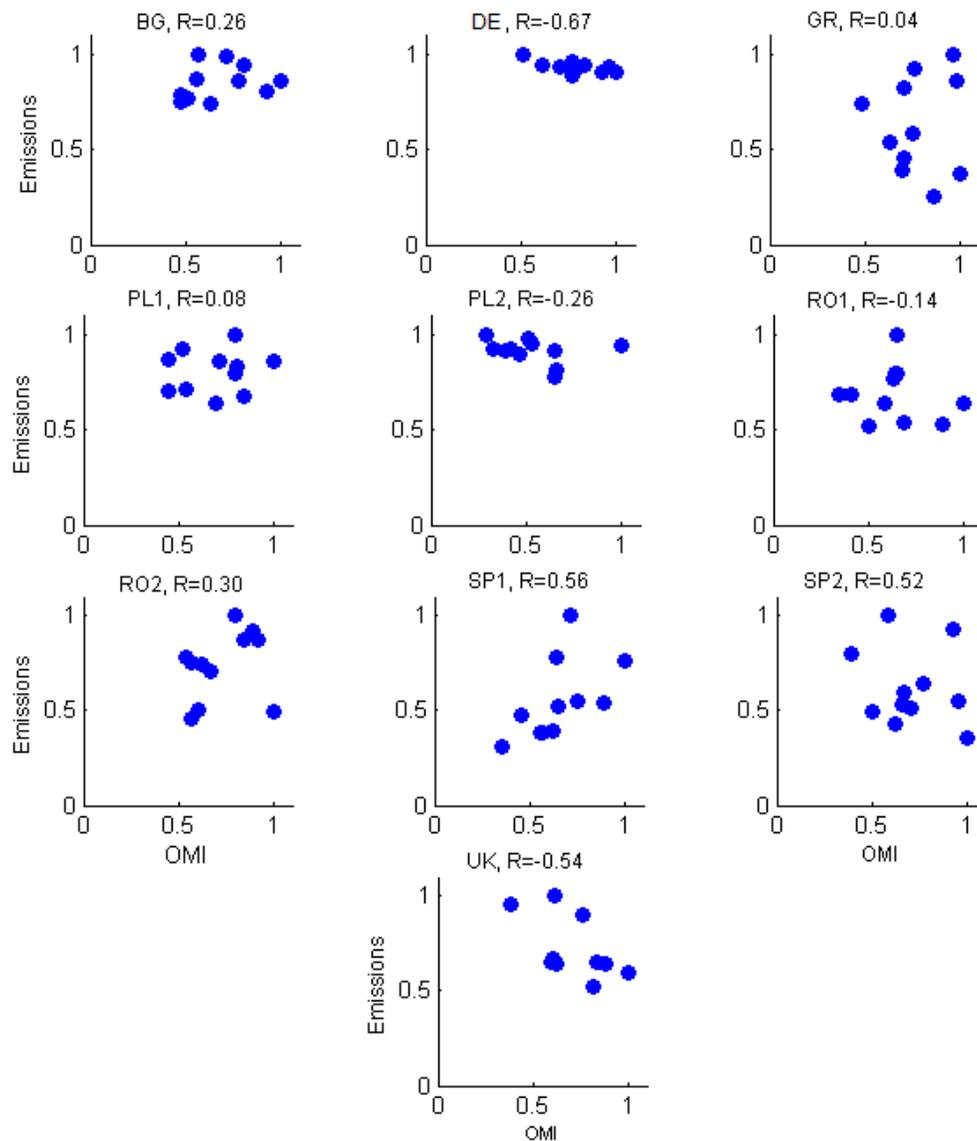


Figure 6. Scatter plots of OMI measurements vs. reported emission for NO₂ (normalized values).

3.2. Sulfur Dioxide

The SO₂ emissions of power plants in Europe have drastically decreased over the last 20 years. Figure 7 shows that during 2005–2015, the observed SO₂ observed by OMI (blue line) has no trend (as in PL1, PL2, SP1, and the UK) or is slightly decreasing (BG, GR, RO1, EO2, and SP2). The emissions (red line), on the other hand, have a clear decreasing trend for most stations, except the UK, with a maximum in 2012, and DE, where the highest emissions were reported also in 2012.

Similar to NO₂, we do not aim to compare satellite measurements, emission, and fuel consumption, but their annual variation (this is also why we use normalized values). The similarities between satellite data and emissions are better for SO₂ than for NO₂. Some discrepancies between the two-time series were seen for the UK, PL1, or DE. Emissions decreased more abruptly than to VCD measurements in GR, PL1, RO2, RO1, SP2, and SP1. Even if the trend was the same, differences between the relative variations of VCDs and emissions were still high (e.g., GR, PL1, RO1, and SP2) (Figure 7). Important factors to be considered are (1) the overpass time of the space sensor over the emission source and, as in the case of NO₂, and (2) the contribution of other sources to the SO₂ amount measured by satellite. Except for the two plants from Spain (SP1 and SP2), SO₂ emissions decreased, on average, after 2010. Satellite measurements confirmed the SO₂ decrease, especially for BG and GR. The descending trend of reported SO₂ emissions was higher compared to the descending trend of SO₂ observed from space. The correlation between the SO₂ content observed from space and the type of fossil fuel may indicate whether the main used fuel is liquid or solid.

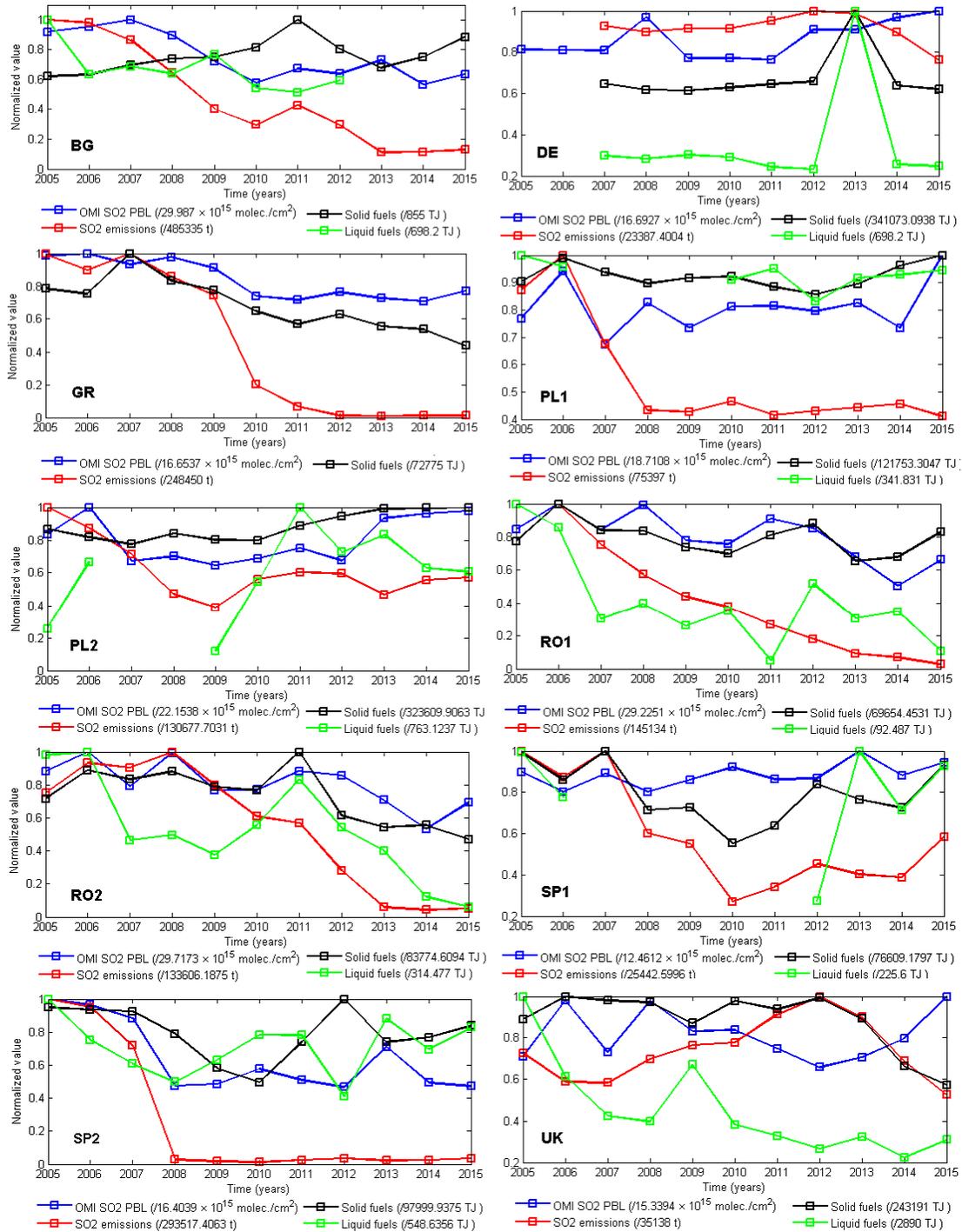


Figure 7. Time series of normalized SO₂ during 2005–2015 using data from satellite instruments (blue) and emissions (red), together with the reported solid fuel (black) and liquid fuel consumption (green).

The large difference between the trend of reported emissions and space observations over the power plants from Greece, Romania, and Spain can be explained by the use of improved SO₂ filter systems, e.g., the case of Romanian LCPs [44,45].

Figure 8 confirms that SO₂ satellite measurements correlated better to the corresponding emissions compared to NO₂ for BG, GR, PL2, RO1, RO2, and SP2. The variation of SO₂ emissions of

the UK, DE, PL1, and SP1 were not supported by satellite measurements of tropospheric column. Similar to the analysis of NO₂, satellite measurements average over a larger area and include background or transport emissions.

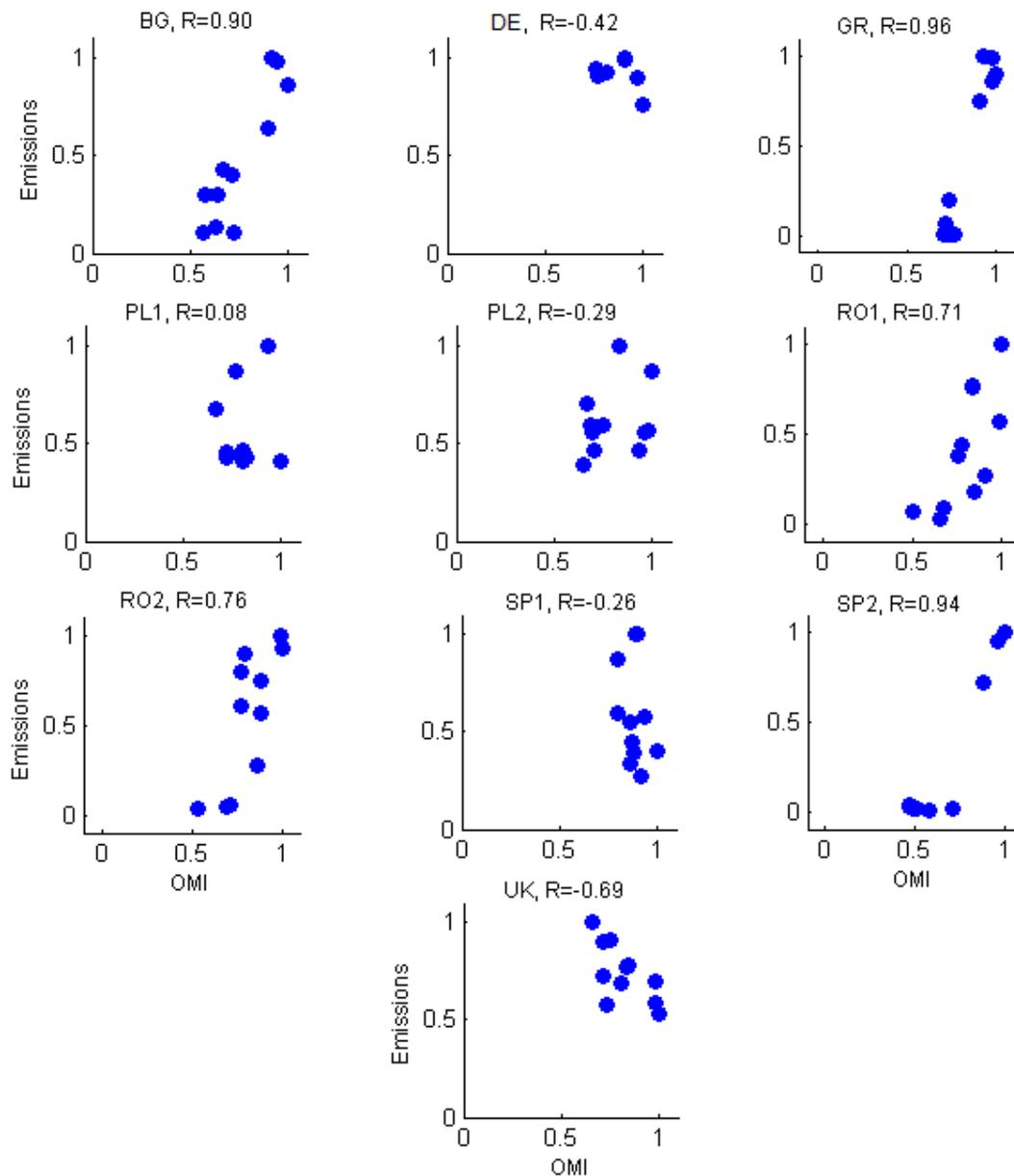


Figure 8. Scatter plots of OMI measurements vs. emissions for SO₂ (normalized values).

A slight decrease was seen in the satellite-based SO₂ for BG, GR, RO1, RO2, and SP2. For the rest (SP1, DE, and the UK), no clear trend could be identified.

A good correlation was at PL2, RO1, RO2, and BG, where at least there were no large discrepancies between the annual variations, and decreased emissions were, most times, supported by accompanying decreases in SO₂ measurements. A good correlation between the two sets was also seen at the DE station, where, despite the small value of the correlation coefficient, both satellite and estimated emissions showed a similar variation of the SO₂ content.

A different variation was seen for emissions at GR, PL1, RO1, and RO2, which went down to 40% of their initial values in 2005, 2006, and 2007. This decrease, however, was not backed up by satellite measurements whose variation is (as for the UK) about 25% of the maximum. Something

marginally similar was seen in SP2 and GR, where the decrease in emissions after 2008 was even more dramatic, going down by almost 90%. Satellite measurements partly supported, in general, the decrease in SO₂, even if the corresponding variation was only 30%–50%. This discrepancy can be explained by possible additional SO₂ loading seized by the satellite instrument. However, natural sources of SO₂ cannot account for the difference, and most SO₂ originated from coal and oil burning, unlike NO₂.

For the UK, the situation is completely different, since both emissions and measurements vary within the same range, but they are anticorrelated. The effects of the cold winter of 2010 and the financial and economic crisis on the SO₂ variability were less visible than for NO₂. This may have been related to the fact that the power plants had improved their SO₂ filtering system, the burning system and the quality of the fuel used, e.g., this would be the case of Romanian power plants, for which the implementation of the EU directives was achieved gradually during 2007–2013 [28]. This was seen both in satellite measurements and in the emissions for stations in Romania (RO1 and RO2).

Some inconsistencies between space observations and ground emissions may also arise from the electricity demand or the energy consumption curve, which is represented by the electricity demand function of time. The space sensor can have the overpass time during high demand of electricity, which corresponds to high emissions. OMI passes over the Romanian LCP around 10–12 UTC, which corresponds to the time interval of high demand for electricity [46]. In such a case, i.e., when the overpass time of the space sensor coincides with a high demand of energy, thus of SO₂, this may explain the fact that the satellite-based SO₂ may overestimates the SO₂ emissions. The reported emissions are based on daily emissions calculations while the space observations are based on one, two, or no observation per day function of orbit. Note that the EU Directives does not impose hourly or daily limits for the emissions, the emission restrictions are quantified annually.

Figures 9 and 10 show the annual changes in emissions relative to 2010, for NO_x and SO₂. We chose 2010 because, by then, the level of national emissions, including those caused by LCPs, should have complied with the NEC Directive (2001/81/EC). According to Figures 5 and 7, emissions for all LCPs presented in this work show descending trends until 2010. After 2010 a clear increasing trend for both NO_x and SO₂ emissions was observed only for power plants in Spain. The NO_x and SO₂ emissions descending trend was the result of the EU Directives, combined with the financial and economic crisis during 2007–2009. Figure 11 shows the influence of the financial and economic crisis on the evolution of the Nominal GDP (Gross Domestic Product) and electrical energy production using fossil fuels during 2005–2015 (Eurostat, <https://ec.europa.eu/eurostat/data/database>).

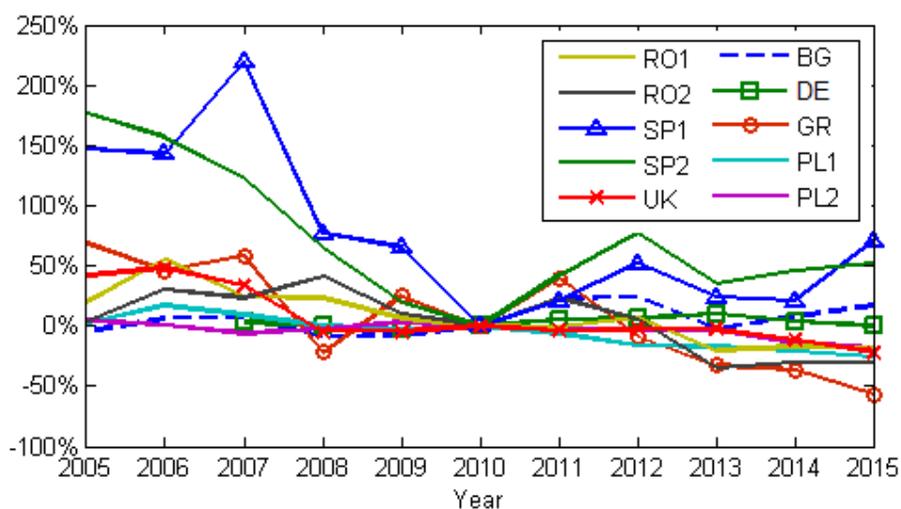


Figure 9. Changes in NO_x emissions relative to 2010, the milestone being NEC Directive.

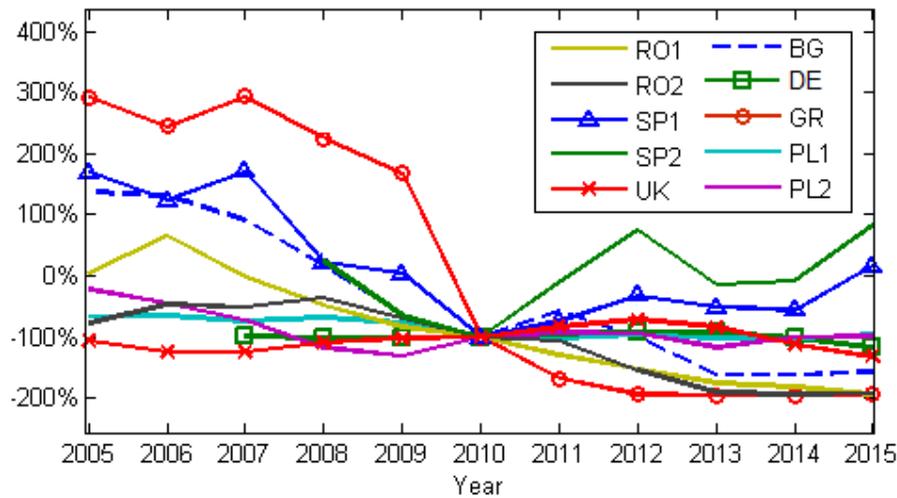


Figure 10. Changes in SO₂ emissions relative to 2010, the milestone being NEC Directive.

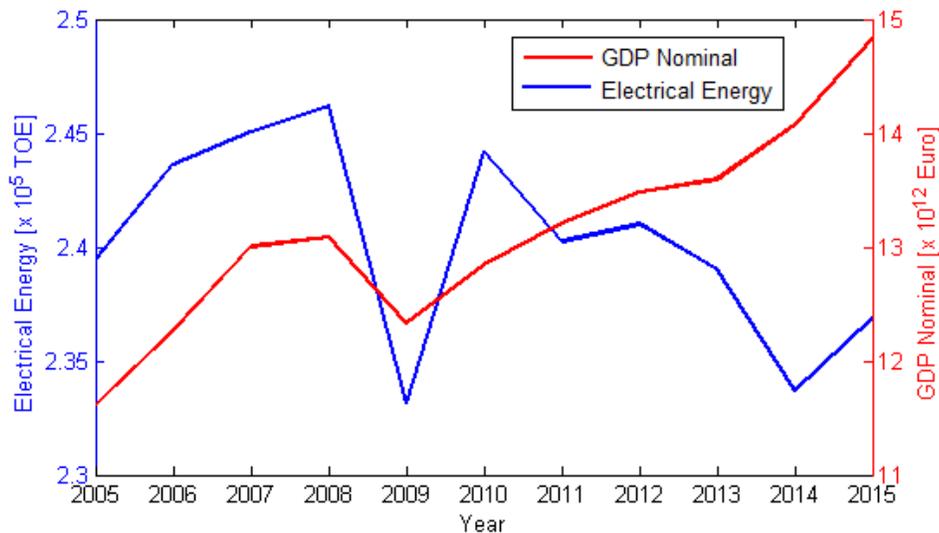


Figure 11. Comparison between Nominal GDP and Electrical Energy over the EU during 2005–2015, where TOE represents a tonne of oil equivalent (source: Eurostat).

4. Conclusions

In this paper, we presented the variation of NO₂ and SO₂ atmospheric emissions attributed to ten very large power plants across Europe (EU-28), quantified by calculated emissions and by satellite observations, during 2005–2015. The main aim of this work was to study the evolution of the NO₂ and SO₂ amount after implementation of the LCP Directive (2001/80/EC), and the NEC Directive (2001/81/EC). Another goal of the comparison was to see to what extent changes in reported emissions to EEA by each country are supported by OMI space-based observations. We presented that OMI observations support the changes in reported emissions to EEA by the EU-28 countries. We identified most of the main drivers for the emissions decrease or increase during 2005–2015. We found that the financial and economic crisis during 2007–2009 had an important role in the emission reduction before 2010.

In general, the NO₂ and SO₂ emissions from LCPs in the Eastern European countries were larger than for Western European countries. During the period between 2005 and 2015, some Eastern Europe socio-political changes triggered by the accession to the European Union took place,

requiring compliance with the legislation imposing pollution limits. These changes were directly reflected in the corresponding emission reductions of NO₂ and SO₂. The reduction in solid fuel consumption (in both Eastern and Western Europe) was balanced by the growth of gas consumption or by renewable and nuclear energy [47]. In Eastern Europe, OMI showed a substantial SO₂ reduction in the proximity of the coal-fired power plants, because flue-gas desulfurization equipment was installed during the study period.

The satellite observations clearly detected a decreasing tendency of NO₂ and SO₂ amount for the entire analyzed period. The satellite observations presented in this paper support the conclusion about the recent decline in NO_x and SO₂ emissions from power plants across Europe, where the EU policies for emission reduction played a key role in the LCP emissions reduction. Future work will focus on the impact assessment of the Industrial Emissions Directive (IED) over the LCPs presented in this study.

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Conflicts of Interest: The authors declare no conflict of interest.

References

1. European Environment Agency. *Air Quality in Europe—2015 Report*; EEA Report No 5/2015; European Environment Agency: Copenhagen, Denmark, 2015.
2. European Environment Agency. *European Union Emission Inventory Report 1990–2016 under the UNECE Convention on Long-Range Transboundary Air Pollution (LRTAP)*; EEA Report 2016, No.6/2018; European Environment Agency: Copenhagen, Denmark, 2018.
3. Maître, A.; Bonnetterre, V.; Huillard, L.; Sabatier, P.; de Gaudemaris, R. Impact of urban atmospheric pollution on coronary disease. *Eur. Heart J.* **2006**, *27*, 2275–2284.
4. Makri, A.; Stilianakis, N.I. Vulnerability to Air Pollution Health Effects. *Int. J. Hyg. Environ. Health* **2008**, *211*, 326–336.
5. Witek, T.J.; Schachter, E.N.; Beck, G.J.; Cain, W.S.; Colice, G.; Leaderer, B.P. Respiratory symptoms associated with sulfur dioxide exposure. *Int. Arch. Occup. Environ. Health* **1985**, *55*, 179–183.
6. Singh, A.; Agrawal, M. Acid rain and its ecological consequences. *J. Environ. Biol.* **2007**, *29*, 15.
7. Rosu, A.; Constantin, D.E.; Georgescu, L. Air pollution level in Europe caused by energy consumption and transportation. *J. Environ. Prot. Ecol.* **2016**, *17*, 1–8.
8. EU. *Directive 2001/80/EC of the European Parliament and of the Council of 23 October 2001 on the Limitation of Emissions of Certain Pollutants into the Air from Large Combustion Plants*; EU: Brussels, Belgium, 2001.
9. EU. *Directive 2008/1/EC of the European Parliament and of the Council of 15 January 2008 Concerning Integrated Pollution Prevention and Control*; EU: Brussels, Belgium, 2008.
10. EU. *Directive 2001/81/EC of the European Parliament and of the Council of 23 October 2001 on National Emission Ceilings for Certain Atmospheric Pollutants*; EU: Brussels, Belgium, 2001.
11. EU. *Directive 2010/75/EU of the European Parliament and of the Council of 24 November 2010 on Industrial Emissions (Integrated Pollution Prevention and Control)*; EU: Brussels, Belgium, 2010.

12. Constantin, D.E.; Merlaud, A.; Voiculescu, M.; Van Roozendaal, M.; Arseni, M.; Rosu, A.; Georgescu, L. NO₂ and SO₂ observations in southeast Europe using mobile DOAS observations. *Carpathian J. Earth Environ. Sci.* **2017**, *12*, 323–328.
13. Yang, J.; Gong, P.; Fu, R.; Zhang, M.; Chen, J.; Liang, S.; Dickinson, R. The role of satellite remote sensing in climate change studies. *Nat. Clim. Chang.* **2013**, *3*, 875.
14. Eisinger, M.; Burrows, J.P. Tropospheric sulfur dioxide observed by the ERS-2 GOME instrument. *Geophys. Res. Lett.* **1998**, *25*, 4177–4180.
15. Burrows, J.P.; Weber, M.; Buchwitz, M.; Rozanov, V.; Ladstätter-Weissenmayer, A.; Richter, A.; DeBeek, R.; Hoogen, R.; Bramstedt, K.; Eichmann, K.-U.; et al. The Global Ozone Monitoring Experiment (GOME): Mission Concept and First Scientific Results. *J. Atmos. Sci.* **1999**, *56*, 151–175.
16. Constantin, D.-E.; Voiculescu, M.; Georgescu, L. Satellite Observations of NO₂ Trend over Romania. *Sci. World J.* **2013**, *2013*, doi:10.1155/2013/261634.
17. Paraschiv, S.; Constantin, D.-E.; Paraschiv, S.-L.; Voiculescu, M. OMI and Ground-Based In-Situ Tropospheric Nitrogen Dioxide Observations over Several Important European Cities during 2005–2014. *Int. J. Environ. Res. Public Health* **2017**, *14*, 1415.
18. Rosu, A.; Rosu, B.; Constantin, D.E.; Arseni, M.; Voiculescu, M.; Georgescu, L.P.; Popa, I. Overview of tropospheric NO₂ using the ozone monitoring observations instrument and human perception about air quality for the most polluting countries across the world. *Carpathian J. Earth Environ. Sci.* **2019**, *14*, 423–430.
19. Levelt, P.; van den Oord, G.; Dobber, M.; Malkki, A.; Visser, H.; de Vries, J.; Stammes, P.; Lundell, J.; Saari, H. The ozone monitoring instrument. *IEEE Trans. Geosci. Remote* **2006**, *44*, 1093–1101.
20. Fioletov, V.E.; McLinden, C.A.; Krotkov, N.; Li, C. Lifetimes and emissions of SO₂ from point sources estimated from OMI. *Geophys. Res. Lett.* **2015**, *42*, 1969–1976, doi:10.1002/2015GL063148.
21. Lin, J.; McElroy, M.B.; Boersma, F. Constraint of anthropogenic NO_x emissions in China from different sectors: A new methodology using multiple satellite retrievals. *Atmos. Chem. Phys. Discuss.* **2010**, *10*, 63–78, doi:10.5194/acp-10-63-2010.
22. Wang, S.; Streets, D.G.; Zhang, Q.; He, K.; Chen, D.; Kang, S.; Wang, Y. Satellite detection and model verification of NO_x emissions from power plants in Northern China. *Environ. Res. Lett.* **2010**, *5*, 044007, doi:10.1088/1748-9326/5/4/044007.
23. Kim, S.-W.; Heckel, A.; McKeen, S.A.; Frost, G.J.; Hsie, E.-Y.; Trainer, M.K.; Grell, G.A. Satellite-observed U.S. power plant NO_x emission reductions and their impact on air quality. *Geophys. Res. Lett.* **2006**, *33*, L22812, doi:10.1029/2006GL027749.
24. van der A., R.J.; Mijling, B.; Ding, J.; Koukouli, M.E.; Liu, F.; Li, Q.; Mao, H.; Theys, N. Cleaning up the air: Effectiveness of air quality policy for SO₂ and NO_x emissions in China. *Atmos. Chem. Phys.* **2017**, *17*, 1775–1789, doi:10.5194/acp-17-1775-2017.
25. Wang, T.; Wang, P.; Theys, N.; Tong, D.; Hendrick, F.; Zhang, Q.; Van Roozendaal, M. Spatial and temporal changes in SO₂ regimes over China in the recent decade and the driving mechanism. *Atmos. Chem. Phys.* **2018**, *18*, 18063–18078, doi:10.5194/acp-18-18063-2018.
26. Russell, A.R.; Valin, L.C.; Cohen, R.C. Trends in OMI NO₂ observations over the United States: Effects of emission control technology and the economic recession. *Atmos. Chem. Phys.* **2012**, *12*, 12197–12209.
27. De Foy, B.; Lu, Z.; Streets, D.G.; Lamsal, L.N.; Duncan, B.N. Estimates of power plant NO_x emissions and lifetimes from OMI NO₂ satellite retrievals. *Atmos. Environ.* **2015**, *116*, 1–11.
28. Krotkov, N.A.; McLinden, C.A.; Li, C.; Lamsal, L.N.; Celarier, E.A.; Marchenko, S.V.; Swartz, W.H.; Bucsel, E.J.; Joiner, J.; Duncan, B.N.; et al. Aura OMI observations of regional SO₂ and NO₂ pollution changes from 2005 to 2015. *Atmos. Chem. Phys.* **2016**, *16*, 4605–4629, doi:10.5194/acp-16-4605-2016.
29. Li, C.; McLinden, C.; Fioletov, V.; Krotkov, N.; Carn, S.; Joiner, J.; Dickerson, R.R. India is overtaking China as the world's largest emitter of anthropogenic sulfur dioxide. *Sci. Rep.* **2017**, *7*, 14304.
30. EU. *Assessing the Effectiveness of EU Policy on Large Combustion Plants in Reducing Air Pollutant Emissions*; EEA Report No 07/2019; EU: Brussels, Belgium, 2019; ISSN 1977-8449, doi:10.2800/13745.
31. EU. *Emissions of Air Pollutants from Large Combustion Plants, Indicator Assessment*; Prod-ID: IND-427-en(a); EU: Brussels, Belgium, 2017; Available online: <https://www.eea.europa.eu/data-and-maps/indicators/emissions-of-air-pollutants-from/assessment-1> (accessed on 15 April 2020).
32. EU. *Emissions of Air Pollutants from Large Combustion Plants in Europe, Indicator Assessment*; Prod-ID: IND-427-en(b); EU: Brussels, Belgium, 2020. Available online:

- <https://www.eea.europa.eu/data-and-maps/indicators/emissions-of-air-pollutants-from-16/assessment> (accessed on 15 April 2020).
33. Singhal, P. Are Emission Performance Standards Effective in Pollution Control? Evidence from the EU's Large Combustion Plant Directive; *DIW Berlin*: Berlin, Germany, 2019; Discussion Paper No. 1773, doi:10.2139/ssrn.3297528. Available online: <https://ssrn.com/abstract=3297528> (accessed on 1 June 2019).
 34. Meyer, A.; Pac, G. Analyzing the characteristics of plants choosing to opt-out of the Large Combustion Plant Directive. *Util. Policy* **2017**, *45*, 61–68.
 35. Boersma, K.F.; Eskes, H.J.; Dirksen, R.J.; van der A, R.J.; Veefkind, J.P.; Stammes, P.; Huijnen, V.; Kleipool, Q.L.; Sneep, M.; Claas, J.; et al. An improved retrieval of tropospheric NO₂ columns from the Ozone Monitoring Instrument. *Atmos. Meas. Tech.* **2011**, *4*, 1905–1928, doi:10.5194/amt-4-1905-2011.
 36. Acker, J.G.; Leptoukh, G. Online Analysis Enhances Use of NASA Earth Science Data. *Eos Trans. AGU* **2007**, *88*, 14–17.
 37. Li, C.; Joiner, J.; Krotkov, N.A.; Bhartia, P.K. A fast and sensitive new satellite SO₂ retrieval algorithm based on principal component analysis: Application to the Ozone Monitoring Instrument. *Geophys. Res. Lett.* **2013**, doi:10.1002/2013GL058134.
 38. Krotkov, N.A.; Carn, S.A.; Krueger, A.J.; Bhartia, P.K.; Yang, K. Band residual difference algorithm for retrieval of SO₂ from the Aura Ozone Monitoring Instrument (OMI). *IEEE Trans. Geosci. Remote Sens.* **2006**, *44*, 1259–1266.
 39. Nickolay, A.; Krotkov, C.L.; Leonard, P. *OMI/Aura Sulfur Dioxide (SO₂) Total Column L3 1 Day Best Pixel in 0.25 Degree x 0.25 Degree V3*; Goddard Earth Sciences Data and Information Services Center (GES DISC): Greenbelt, MD, USA, 2015; doi:10.5067/Aura/OMI/DATA3008.
 40. Cattiaux, J.; Vautard, R.; Cassou, C.; Yiou, P.; Masson-Delmotte, V.; Codron, F. Winter 2010 in Europe: A cold extreme in a warming climate. *Geophys. Res. Lett.* **2010**, *37*, doi:10.1029/2010gl044613.
 41. Castellanos, P.; Boersma, K.F. Reductions in nitrogen oxides over Europe driven by environmental policy and economic recession. *Sci. Rep.* **2012**, *2*, 1–7, doi:10.1038/srep00265.
 42. Liu, F.; Zhang, Q.; Zheng, B.; Tong, D.; Yan, L.; Zheng, Y.; He, K. Recent reduction in NO_x emissions over China: Synthesis of satellite observations and emission inventories. *Environ. Res. Lett.* **2016**, *11*, 114002.
 43. Vrekoussis, M.; Richter, A.; Hilboll, A.; Burrows, J.P.; Gerasopoulos, E.; Lelieveld, J.; Barrie, L.; Zerefos, C.; Mihalopoulos, N. Economic crisis detected from space: Air quality observations over Athens/Greece. *Geophys. Res. Lett.* **2013**, *40*, 458–463, doi:10.1002/grl.50118.
 44. ANPM. Air Quality Management Program of Gorj County 2010–2013. Available online: www.anpm.ro/anpm_resources/migrated_content/uploads/20817_Program%20CA%20jud_Gorj.pdf (accessed on 12 June 2017).
 45. Merlaud, A.; Belegante, L.; Constantin, D.-E.; Den Hoed, M.; Meier, A.C.; Allaart, M.; Ardelean, M.; Arseni, M.; Bösch, T.; Brenot, H.; et al. The Airborne Romanian Measurements of Aerosols and Trace gases (AROMAT) campaigns. *Atmos. Meas. Tech. Discuss.* **2020**, doi:10.5194/amt-2019-496. in review.
 46. Bogdan, M.; Stanescu, D.; Stanescu, C.; Achim, N. Aspecte didactice privind contorul inteligent de energie electrică. In *Proceedings of the Simpozionul International "Contorizare Inteligenta"*, Sibiu, Romania, 14–17 November 2017.
 47. Vestreng, V.; Myhre, G.; Fagerli, H.; Reis, S.; Tarrasón, L. Twenty-five years of continuous sulphur dioxide emission reduction in Europe. *Atmos. Chem. Phys.* **2007**, *7*, 3663–3681, doi:10.5194/acp-7-3663-2007.

